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17	A $\delta^{15}N$ assessment of nitrogen deposition for the endangered epiphytic orchid <i>Laelia speciosa</i>
18	from a city and an oak forest in Mexico
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ABSTRACT

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Atmospheric nitrogen deposition poses a major threat to global biodiversity. Tropical epiphytic plants are especially at risk given their reliance on atmospheric sources of nutrients. The leaf, pseudobulb, and root carbon and nitrogen content, C:N ratio, as well as the nitrogen isotopic composition were studied for individuals of Laelia speciosa from a city and from an oak forest in Mexico. The nitrogen content of leaves was similar between the city and the oak forest, reaching $1.3 \pm 0.2\%$ (dry mass). The $\delta^{15}N$ of leaves, pseudobulbs, and roots reached $5.6 \pm 0.2\%$ in the city, values found in sites exposed to industrial and vehicular activities. The $\delta^{15}N$ for plant from the oak forest amounted to $-3.1 \pm 0.3\%$, which is similar to values measured from sites with low industrial activities. Some orchids such as Laelia speciosa produce a single pseudobulb per year, i.e., a water and nutrient storage organ, so the interannual nitrogen deposition was studied by considering the ten most recent pseudobulbs for plants from either site formed between 2003 and 2012. The C:N ratio of the of ten most recent pseudobulbs from the oak forest, as well as that of the pseudobulbs formed before 2010 for plants in the city were indistinguishable from each other, averaging $132.4 \pm$ 6.5, while it was lower for the two most recent pseudobulbs in the city. The δ^{15} N values of pseudobulbs from the oak forest averaged $-4.4 \pm 0.1\%$ for the entire series. The δ^{15} N ranged from $0.1 \pm 1.6\%$ for the oldest pseudobulb to 4.7 \pm 0.2% for the pseudobulb formed in 2008 onwards. Isotopic analysis and the C:N ratio for L. speciosa revealed that rates of nitrogen deposition were higher in the city than in the forest. The $\delta^{15}N$ values of series of pseudobulbs showed that it is possible to track nitrogen deposition over multiple years.

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Key words: Atmospheric pollution; Conservation Physiology; global change; industrial activities; Neo-Volcanic

61 Axis.

Introduction

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Atmospheric nitrogen deposition has more than doubled since the mid-20th century as a result of an increased release of reactive nitrogen species from industrial and agricultural origin (Phoenix et al. 2006; Galloway et al. 2008). This important component of global change has driven a loss of biodiversity, which has been widely documented in Europe, the USA, and China (Xue-Yan et al. 2008; Bobbink et al. 2010). However, studies are scant for megadiverse countries like Mexico, which tend to have developing economies and ongoing industrialization processes, making it urgent to determine the effects of nitrogen deposition on their biodiversities (Austin et al. 2013). In this respect, an interest on tracing sources of deposition has increased and it can be studied by means of the isotopic composition of the vegetation.

Atmospheric nitrogen deposition confers a distinctive isotopic signature to vegetation. The isotopic composition of the reactive nitrogen species (NOx and NHx) that result from the atmospheric N₂ rupture can be either positive or negative (Heaton 1990). In particular, the NOx from electric plants, coal, gasoline, diesel, and trash burning has δ^{15} N between +3.7 and +26% (Widory 2007). However, these isotopic values can also be negative, ranging from -13 to -2% (Redling et al. 2013). Whether the values are negative or positive will depend on the type of material burnt, the nitrogen in the fuel, the various isotopic fractionations associated with the thermal production of NOx, when the fuel goes through the engine and is mixed with air O2 and N2 during the combustion (Liu et al 2012; Felix et al. 2013; Felix and Elliot 2014). The δ^{15} N of biogenic emissions (NHx) of soils in rural areas are negative and can fluctuate between -50 and -20%, including, organic wastes of animals and volatilization of nitrogen from agriculture (Hoering 1957; Li and Wang 2008; Felix et al. 2012). Once emitted the NOx or the NHx react with various compounds in the atmosphere, such reactions involve isotopic fractionations, that result in compounds such as NO_3^- and NH_4^+ , with different $\delta^{15}N$ values in rain ranging from -15 to +15%, the NH_4^+ is more negative than the NO₃⁻ (Heaton1990; Xiao and Liu 2002; Heaton et al. 2004; Elliot et al. 2009; Xiao et al. 2012). These isotopic values directly affect the ¹⁵N composition of plants of different sites. In particular, plants exposed to industrial and vehicular emissions tend to be ^{15}N -enriched. For example, the $\delta^{15}N$ of mosses growing close to busy urban roads range between -1 and 6%, while plants from less polluted areas tend to be ¹⁵N depleted as is the case for mosses in rural areas, whose δ^{15} N values range from -2 to -12% (Ammann et al. 1999; Pearson et al., 2000; Stewart et al., 2002; Liu et al. 2007; Power and Collins 2010). However, in cities where the industrial activities

release NHx, the δ^{15} N of plants are more negative than in the field (Stewart et al. 2002; Liu et al. 2008; Xiao et al. 2010).

A lifeform of particular interest for studying the effects of deposition is that of epiphytes, which account for up to 50% of vascular plant species in some tropical ecosystems and are particularly susceptible to atmospheric nitrogen deposition, given their reliance on atmospheric sources of nutrients (Mondragón et al. 2015). These plants, whose roots have no contact with the forest soil, where the nitrogen fractionation takes place, reflect the δ^{15} N of the prevalent atmospheric nitrogen deposition (Delwiche and Steyn 1970; Hietz et al. 1999; Stewart et al. 2002; Zotz et al. 2010).

Laelia speciosa (Kunth) Schltr (Orchidaceae) is an endangered endemic epiphytic orchid from central Mexico with significant cultural importance (Halbinger and Soto-Arenas 1997). Vast numbers of individuals are illegally extracted and sold in streets and markets during May and June owing to their colorful flowers. This has caused a significant reduction of natural populations that has led to its special protection status by the environmental authority (Soto-Arenas and Solano-Gómez 2007; NOM-059-SEMARNAT-2010; Mondragón et al. 2015).

Moreover, oak forests, the natural habitat for this orchid, are highly threatened by changes in land use and climate (Villers-Ruiz and Trejo-Vázquez 2000; Soto-Arenas and Solano-Gómez 2007; Rehfeldt et al. 2012; Gudiño et al. 2015).

Under increasing rates of nitrogen deposition of up to 20 kg N ha⁻¹ y⁻¹, the carbon and nitrogen content increase and the C:N ratio decreases in epiphytes responding to increased photosynthetic rates (Díaz-Álvarez et al. 2015). But higher doses of nitrogen can produce noxious physiological effects on epiphytes, given their direct exposure to the atmosphere (Hietz et al. 1999; Zotz et al. 2010). For instance, higher nitrogen causes the reduction of chlorophyll content and chlorophyll fluorescence, as well as the reduction of leaf and pseudobulb production for *Laelia speciosa*, for which increased tissue nitrogen content is concurrent with nitrogen isotopic fractionation (Díaz-Álvarez et al. 2015). A plant's ability to accumulate foliar mass is limited by large quantities of nitrogen, because the availability of other nutrients is limited. For instance, excessive availability of nitrogen results in an imbalance of the nitrogen to magnesium ratio in the leaf because the cell releases protons (H⁺), which lowers the pH and inhibits chlorophyll production causing the loss of Mg²⁺ (Mangosá and Berger, 1997; Sánchez et al. 2000; Nakaji et al. 2001; Britto and Kronzucker, 2002; Wortman et al. 2012; Díaz-Álvarez et al. 2015). It is thus important to characterize the nitrogen deposition in the natural habitat of this species.

Environmental changes, such as variation in air temperature and available water, have been studied over multiple years by determining the natural abundance of some stable isotopes in the vegetation. For instance, H, C, and O isotopes of cactus thorns and tree rings indicate the occurrence of droughts in the Sonoran Desert and temperature and precipitation variations in Europe (West et al. 2006; English et al. 2007; 2010; Cufar et al. 2014). Also, changes in nitrogen deposition have been determined from δ^{15} N of tree rings in the tropics (Hietz et al. 2011; Van Der Sleen et al. 2015). In this respect, some orchids, including *L. speciosa*, produce one pseudobulb per year, which remains attached to the plant for several years (Soto-Arenas and Solano-Gómez 2007). Thus, these storage organs of water, C, and other nutrients could be useful to reconstruct the interannual variation of nitrogen deposition over the plant's lifespan (Dressler 1981; Ng and Hew 2000; Soto-Arenas and Solano-Gómez 2007).

Leaf, pseudobulb, and root C and N content, C:N ratio, as well as the natural abundance of ¹⁵N were determined for individuals of *L. speciosa* growing in a city and growing in an oak forest in order to compare the signal that nitrogen deposition had on the tissue composition of this orchid at each site. In addition, series of annually produced pseudobulbs were analyzed to determine whether this orchid can record the historical changes on nitrogen deposition at either site.

Materials and methods

Study sites

The effect that atmospheric deposition had on tissue nitrogen composition was evaluated for the orchid *Laelia speciosa* from two contrasting sites in central Mexico. The first site was the Instituto de Investigaciones en Ecosistemas y Sustentabilidad, Universidad Nacional Autónoma de México (19° 38' 55.9" N; 101° 13' 45" W), located in the city of Morelia whose population is 800,000 habitants. Here, the mean annual temperature and annual precipitation are 18.3 °C and 773 mm, respectively, and dominant winds blow from the southwest and northwest (Servicio Meteorológico Nacional 2012; Instituto Nacional de Estadística y Geografía 2013). Orchids that had been rescued from a road construction site in 2004 were kept on the original oak branches of *Quercus deserticola* Trel. (Fagaceae) from which they were collected and placed inside a shadehouse. This shadehouse consists of a metallic structure surrounded with a shade-cloth (85% transmitance), including the top and sides, which allows the free movement of water and gases, while simulating a canopy leading to a temperature difference of up to 4 °C below ambient at midday.

The second site was one of the last natural refuges of *Laelia speciosa* in Mexico, an oak forest that is located at the Cerro el Olvido, Tzintzuntzán, Michoacán (19° 37' 59" N, 101° 29' 09" W, 2361 m; García-Cruz et al. 2003). The dominant species at this site is *Quercus deserticola* which is the most common phorophyte for *L. speciosa* (Soto-Arenas 1994). Oak trees in this forest reach 10 m in height and grow in rocky, clay or clay-loam soils (unpublished observations). The mean annual temperature at Cerro el Olvido is 16.1 °C and the annual precipitation is 758 mm, with a rainy season from May to September, and the dominant winds blow from the southwest and the northwest (Servicio Meteorológico Nacional 2012).

Characterization of the atmospheric nitrogen deposition

Given that no infrastructure exists for measuring atmospheric pollution in the study region, the prevalent nitrogen deposition at each site was assessed by means of mosses. Indeed, mosses have been broadly utilized as bioindicators of atmospheric nitrogen deposition because their nitrogen content is correlated with the rates of deposition and because their isotopic fractionation during assimilation is nill, reflecting the δ^{15} N values of the prevalent atmospheric deposition (Bragazza et al. 2005; Zechmeister et al. 2008; Wilson et al. 2009; Power and Collins 2010). In particular, the genus *Braunia*, which is widely distributed in various Mexican ecosystems including urban environments, has shown to be an adequate bioindicator of atmospheric nitrogen deposition (Arciga-Pedraza 2009; Cardenas and Delgadillo 2009). In order to determine the δ^{15} N values of the atmospheric nitrogen deposition as well as to determine the differences in rates of deposition by means of the nitrogen content of the mosses, samples of the moss *Braunia* sp. growing on the same oak branches that the orchids and less than 15 cm away of them in the oak forest and in the city were collected. In addition, to preclude any contamination from organic material from the phorophyte, i.e., the so called canopy soil, moss samples were also collected from exposed granite rock in the oak forest or exposed concrete in the city. The moss samples were analyzed as described below.

Plant material

Laelia speciosa (Kunth) Schltr is a sympodial epiphytic orchid with big and showy flowers, pink to lilac-purple petals, and a white lip. Flowers are produced during the spring. This orchid grows in sub-humid temperate climates, between 1250 and 2500 m (Soto-Arenas and Solano-Gómez 2007). The most recent leaf, pseudobulb, and root were collected from 4 individuals at each site, whose leaf and root length ranged from 8 to 12 cm and 5 to 10 cm

respectively. In the oak forest, orchids were collected from the outer portion of trees where canopy soil cannot be formed given a lack of plant cover. Indeed, these plants were not in contact with decomposing plant material and were at a minimum height of 5 m above the ground and a minimum radial length of 3 m from the stem of *Q*. *deserticola*.

Interannual atmospheric nitrogen deposition

In order to assess whether the nutrient-storing pseudobulbs of *L. speciosa* can record the changes of atmospheric nitrogen deposition over multiple years, series of consecutive pseudobulbs comprised by the ten most recent organs were collected for 4 individuals from each site (Fig. 1) for conducting the C and N analyses described below.

Carbon and nitrogen content and isotopic $\delta^{15}N$

- Plant material was dried at 80 $^{\circ}$ C in a gravity convection oven until reaching constant weight (Drennan 2009; Díaz-Álvarez et al. 2015). The dried plant samples were ground to a fine powder in a ball mill (Retsch MM300; Retsch, Vienna, Austria), wrapped into tin capsules (Costech Analytical, Inc. Valencia, California, USA), and weighed with a microbalance (0.01 mg, Sartorius, Göttingen, Germany). For each sample, the carbon and nitrogen content, as well as their δ^{15} N values were determined at the Stable Isotope Facility, University of Wyoming (Laramie, Wyoming, USA), with a Carlo Erba EA 1110 elemental analyzer (Costech Analytical Inc., Valencia, CA, USA) attached to a continuous flow isotope ratio mass spectrometer (Finnigan Delta Plus XP, Thermo Electron Corp, Waltham, MA). Nitrogen isotope ratios, reported in parts per thousand, were calculated relative to atmospheric air standards. The
- natural abundances of ¹⁵N were calculated as:
- $\delta^{15} N_{\text{(\% versus air)}} = (R_{\text{sample}}/R_{\text{standard}} 1) \times 1000$
- where, R_{sample} and $R_{standard}$ is the ratio of $^{15}N/^{14}N$ for the sample and standard respectively. (Ehleringer and Osmond 1989; Evans et al. 1996). The analytical precision for $\delta^{15}N$ was \pm 0.06% (SD).

Statistical analyses

- Data normality and homoscedasticity were confirmed before conducting the following statistical analyses.
- Atmospheric nitrogen deposition, characterized by means of the moss *Braunia* sp. from the two sites, was analyzed
- with a one-way ANOVA followed by pairwise Holm-Sidak post-hoc tests (P < 0.05). The effects of the site of

origin (two levels) on the nitrogen content, C:N ratio, and δ^{15} N in *L. speciosa* organs (three levels) were analyzed by means of a two-way ANOVA followed by pairwise Holm-Sidak post-hoc tests (P < 0.05). In turn, differences between the series of 10 consecutive pseudobulbs from two different sites were analyzed by means of a nested ANOVA followed by Holm-Sidak post-hoc tests (P < 0.05). All analyses were conducted on R (version 2.12.2, R Core Team R Foundation for Statistical Computing, Vienna, Austria).

Results

Characterization of the atmospheric nitrogen deposition

The nitrogen content of the mosses averaged $1.7 \pm 0.1\%$ (dry mass basis) regardless of the site (Tables 1, 2). Their C:N ratio averaged 21.5 ± 0.8 and no significant difference was found between sites. The δ^{15} N reached $3.3 \pm 0.1\%$ in the city, contrasting with the $-4.6 \pm 0.9\%$ measured in the oak forest (P < 0.05; Fig. 2).

Carbon and nitrogen content and δ¹⁵N for *Laelia speciosa*

The carbon content of leaves, pseudobulbs, and roots of *Laelia speciosa* in the city reached 44.4 ± 0.3 % (dry mass basis), it was 8.5% lower than for orchids in the oak forest (P < 0.05; Tables 1,2). In the city, the carbon content of pseudobulbs was 43.4 ± 0.2 % and it was 10% lower than for the roots (P < 0.05). No significant differences were found between leaves and pseudobulbs (P = 0.12) nor between leaves and roots (P = 0.31). In the oak forest, the carbon content was similar among organs averaging 45.4 ± 0.2 % ($P \ge 0.05$).

The nitrogen content of pseudobulbs and roots in the city averaged $0.9 \pm 0.1\%$ (dry mass basis), and it was lower than for the leaves that reached $1.4 \pm 0.1\%$ (P < 0.05; Fig 3). The nitrogen content of leaves, pseudobulbs, and roots was different among each other in the oak forest (P < 0.05). The nitrogen content of leaves in the oak forest was similar to the nitrogen content of the leaves in the city and averaged $1.3 \pm 0.2\%$ (Fig 3; Table 1). Similar was the case for pseudobulbs and roots whose nitrogen content averaged 0.6 ± 0.1 and $0.9 \pm 0.1\%$ respectively for both sites (Table 2).

The C:N ratio was similar among the organs of orchids from the city, where it averaged 42.1 ± 1.5 . In turn, the C:N ratio for the organs of orchids from the oak forest was different between pseudobulbs and leaves (P < 0.01) and between pseudobulbs and roots (P = 0.02). No significant differences were found between leaves and roots, whose C:N ratio averaged 49.7 ± 0.5 ($P \ge 0.05$). The C:N ratio of orchids growing at the oak forest was generally

the same as that of orchids growing in the city. Indeed, a C:N ratio of 39.1 ± 2.5 was found for leaves of orchids from both sites, while it averaged 50.0 ± 2.0 for the roots. An exception was observed for the C:N ratio of 127.6 ± 16.34 for pseudobulbs at the oak forest that was 62% higher than for the pseudobulbs in the city (P < 0.001; Fig. 2; Table 1).

The δ^{15} N values were positive for plants from the city and negative for plants from the oak forest (P < 0.001). For the plants in the city, differences were found between leaves and roots (P = 0.01) and between pseudobulbs and roots (P = 0.02). However, no differences were found between leaves and pseudobulbs, averaging $5.6 \pm 0.01\%$. For the orchids from the oak forest, δ^{15} N averaged $-3.9 \pm 0.4\%$ and no significant differences were found among organs (Table 1; Fig. 2)

Interannual atmospheric nitrogen deposition

The C:N ratios for the entire series of pseudobulbs of the plants from the oak forest, as well as the C:N ratios for the pseudobulbs formed in 2010 and before in the city were similar and averaged 132.4 ± 6.5 (Fig. 4, Table 3). In contrast, for the two most recent pseudobulbs that grew in the city the C:N ratios halved (P < 0.05).

The δ^{15} N values of pseudobulbs from the oak forest were constant over one decade and averaged $-4.4 \pm 0.1\%$ for the entire series ($P \ge 0.05$; Fig 5). In contrast, the δ^{15} N values of pseudobulbs from the city showed a gradual increase from $0.1 \pm 1.6\%$, for the oldest pseudobulb formed at the rescue site in 2003, until a plateau was reached at $4.8 \pm 0.3\%$ for pseudobulbs formed 2007 onwards.

Discussion

Plants tend to increase their photosynthetic rates as nitrogen becomes more available and, in consequence, they accumulate more carbon (Brown et al. 1996; Shangguan et al. 2000; Bauer et al. 2004; Le Bauer and Treseder 2008; Díaz-Álvarez et al. 2015). For example, when *L. speciosa* is exposed to simulated nitrogen deposition, its leaf carbon content peaks at 46% of the dry mass under 20 kg of N ha⁻¹ year⁻¹, nearly 5% higher than for plants exposed to 2.5 kg of N ha⁻¹ year⁻¹ (Díaz-Álvarez et al. 2015). In the present study, the carbon content of the bioindicator *Braunia* sp. and of the orchids, which was the same at both sites, suggest that nitrogen deposition is not high enough to induce significant changes in CO₂ fixation nor changes in the tissue carbon content

In the present study, the nitrogen content for the moss from both sites was consistent with values measured under low deposition rates, which usually amounts to 1.5% and up to 2.4% in some monocots (Chapman 1965; Mills and Jones 1996; Epstein 1972; 1999). The leaves of *L. speciosa* from both sites also had similar nitrogen content with their values being similar to those of succulent leaves of several species of epiphytic orchids from low nitrogen deposition environments (Hietz et al. 1999; Wania et al. 2002; Cardelús and Mack 2010; Mardegan et al. 2011). The nitrogen content for the plants in this study was substantially lower than for plants from cities with high rates of nitrogen deposition such as London or Mexico City (Power and Collins 2010; Arciga-Pedraza 2010).

While carbon and nitrogen content increase at different rate in response to nitrogen availability, the C:N ratio for the moss and for the leaves and roots of *L. speciosa* was similar at both sites. For the case of the orchid, this parameter was close to the ratio of 56 measured for orchids from the Central Brazilian Amazon, a clean environment (Mardegan et al. 2011). Contrasting with the lack of response from leaves and roots, the C:N ratio of pseudobulbs was different between the sites, which could be the result of a slightly higher accumulation of nitrogen in the city. The lowest nitrogen content of 0.6% was measured for pseudobulbs in the oak forest because these organs predominantly store carbohydrates, rather than minerals (Davidson 1960; Hew and Ng 1996; Ng and Hew 2000).

Plants that grow close to sources of pollution reflect the isotopic signals of the pollutants in their tissues (Stewart et al. 1995; Ammann et al. 1999; Stewart et al. 2002; Liu et al. 2008; Power and Collins 2010; Redling et al. 2013). For instance, negative isotopic values found in compounds derived from NHx (agriculture, animal manure, and sometimes vehicle exhaust) result from secondary compounds formed during reactions after emission (Redling et al. 2013). In contrast, compounds derived from the NOx that are released from industrial activities and fossil fuel burning have positive $\delta^{15}N$ (Xiao and Liu 2002; Xiao et al. 2012). For example, $\delta^{15}N$ of -11% are found for epiphytic plants growing close to a city in Brazil where the main nitrogenous pollutants are NHx (-41%), meanwhile $\delta^{15}N$ of plants from the field are less negative (-3.1%; Stewart et al. 2002). For epiphytes from a Costa Rican cloud forest (low nitrogen deposition) the $\delta^{15}N$ reaches -4% (Hietz et al. 2002; Wania et al. 2002). In contrast, the $\delta^{15}N$ for spruce needles from Norway decreases from 2% in the proximity of a highway where nitrogen deposition is dominated by NOx to -3% at 1 km into the forest (Ammann et al. 1999). The negative $\delta^{15}N$ measured in the present study for the mosses and the orchids from the oak forest suggest that these plants assimilated nitrogen from a clean environment (low nitrogen deposition). Indeed, the total emissions in the municipality where the oak forest is located amounted to a 477 ton NOx and 73 ton NH₃ in 2008 (Sub-Sistema del Inventario Nacional de

Emisiones a la Atmósfera de México 2016). In contrast, positive δ^{15} N were observed for both the moss and the orchid growing in the city, where the emissions were substantially higher than in the forest in 2008, reaching 57,081 ton NOx and 2,732 ton NH₃ (Sub-Sistema del Inventario Nacional de Emisiones a la Atmósfera de México 2016).

The fact that the δ^{15} N was similar among the different organs indicates that isotopic fractionation during the translocation process was very low or absent in plants exposed to the current rates of atmospheric nitrogen deposition at either the forest or the city. These observations contrast with an isotopic fractionation of 3‰ determined for *L. speciosa* growing under a nitrogen deposition of 80 kg N ha⁻¹ year⁻¹ (Díaz-Álvarez et al. 2015).

Physiological responses to climate changes over the lifespan of a plant have been studied with tree rings and cactus spines (English et al. 2007; Hietz et al. 2011; Van Der Sleen et al. 2015). Here, pseudobulb chronosequences developed over a decade effectively recorded the interannual nitrogen deposition at the two sites considered in this study. The C:N ratio observed for the series of pseudobulbs was similar between sites until 2010, when a significant decrease became apparent for the plants in the city. Presumably, this is the result of 8 years of exposure to urban nitrogen deposition. However, it was not enough to decrease the C:N ratio of pseudobulbs before 2011, suggesting that the rates of nitrogen deposition in the city were only slightly higher than in the oak forest. The δ^{15} N values confirmed that the pseudobulbs formed in 2004 and before grew at the rescue site, whereas the more recent pseudobulbs were formed in the city. It was interesting to observe that the first few pseudobulbs formed in city did not show a typical urban δ^{15} N, but apparently the stored nitrogen that had been assimilated in the rescue site was translocated from the older pseudobulb developed in the city. Over time, the nitrogen from the rescue site was depleted, leaving the urban deposition as the only source of nitrogen for the ensuing pseudobulbs, as evidenced by the positive δ^{15} N values of the six most recent pseudobulbs.

Isotopic analyses for *L. speciosa* enabled the distinction between nitrogen deposition in the city and the oak forest. Also, the C:N ratio of the series of pseudobulbs showed that deposition was already slightly higher in the city than in the oak forest, but not enough to effect physiological damage. In turn, the δ^{15} N values of the series of pseudobulbs showed that it is indeed possible to track nitrogen deposition and detect possible changes over various years. In addition to land use and climate changes, the risk for *L. speciosa* will increase as the atmospheric nitrogen deposition increases above 20 kg N ha year $^{-1}$ (Díaz-Álvarez et al. 2015). This is already happening in some regions of Mexico City, where nitrogen deposition rates reach 48 kg N ha $^{-1}$ year $^{-1}$ (Secretaría del Medio Ambiente del Distrito Federal 2012). Granted that these alarmingly high rates of nitrogen deposition do not overlap with the

distribution range of L. speciosa, but deposition rates will most likely exceed 25 kg N ha year ⁻¹ by mid-century in the regions of central Mexico where Laelia speciosa is endemic (Galloway et al. 2004; Phoenix et al. 2006; Galloway et al. 2008). Thus, L. speciosa could disappear in the following decades if necessary actions are not taken to reduce the increasing rates of nitrogen deposition in the regions where this epiphytic orchid still grows. Acknowledgements

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Table 1. Two-way ANOVA for nitrogen content, carbon content, C:N ratio and δ^{15} N of *Laelia speciosa* growing in the city and the oak forest.

		Nitrogen content		Carbon content		C:N Ratio		$\delta^{15}N$	
	d.f.	F	P	F	P	\overline{F}	p	\overline{F}	p
Organ	2	4.07	0.001	5.75	0.01	21.1	0.001	828	0.02
Site	1	20.6	0.06	6.40	0.02	27.0	0.001	4.59	0.001
Organ × Site	2	0.88	0.43	0.10	0.90	12.9	0.001	3.74	0.04

Table 2. Carbon and nitrogen content (dry mass basis) for plant tissue from an oak forest and from a city in Michoacán, Mexico. Data are shown as mean \pm 1 s.e. (n = 4 individuals). For each element, different letters indicate statistically significant differences from pairwise Holm-Sidak tests (P < 0.05).

	Oak forest		City			
	Carbon	Carbon Nitrogen		Nitrogen		
	(% dry mass)	(% dry mass)	(% dry mass)	(% dry mass)		
Braunia sp.						
Inert substrate	39.1 ± 0.24 a	1.57 ± 0.03 c	$37.6 \pm 0.32 a$	1.62 ± 0.09 c		
Epiphytic	35.4 ± 0.24 a	1.66 ± 0.34 c	38.6 ± 0.64 a	2.00 ± 0.05 c		
Laelia speciosa						
Leaves	$45.6 \pm 0.20 \ b$	$1.3 \pm 0.30 \ d$	$44.5 \pm 0.4 \text{ b}$	$1.3 \pm 0.10 \text{ d}$		
Pseudobulbs	$44.6 \pm 0.30 \ b$	0.4 ± 0.10 e	$43.4 \pm 0.2 \ b$	$0.9 \pm 0.10 \text{ f}$		
Roots	$46.0 \pm 0.10 \text{ b}$	$0.8 \pm 0.10 \text{ f}$	$45.2 \pm 0.2 \text{ b}$	$0.9 \pm 0.10 \text{ f}$		

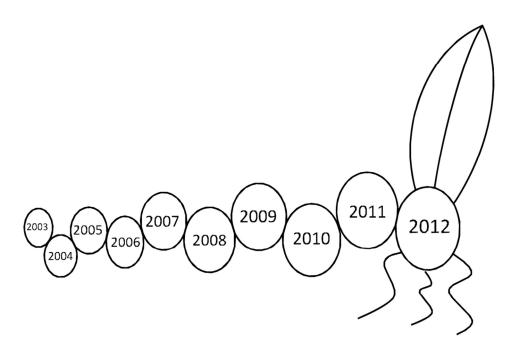
Table 3. Nested ANOVA for the elemental and isotopic composition for series of ten consecutive pseudobulbs of L. speciosa from the city and the oak forest.

554						
555			C:N ratio		δ^{15} N	
556		d.f.	\overline{F}	P	\overline{F}	P
557	Van					
558	Year	1	7.30	0.001	4.13	0.045
559	Year × site	1	0.69	0.405	99.8	1.49×10^{-5}
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574 Figure 1. Morphology of Laelia speciosa showing the series of ten consecutive annually produced pseudobulbs. The 575 most recent pseudobulb has been formed during the past year's growing season. The number inside each pseudobulb 576 indicates the year it was formed. 577 Figure 2. C:N ratio and δ^{15} N for *Laelia speciosa* organs and mosses growing near the orchid on the oak branches 578 579 and mosses from granite or concrete from two sites in Michoacán, Mexico. Data are shown as mean \pm S.E. (n = 4 580 individuals per site). The arrows indicate the mosses growing on exposed bare rock at the oak forest or exposed on 581 concrete at the city. 582 583 Figure 3. Nitrogen content (dry mass basis) for the organs of L. speciosa from two sites in Michoacán, Mexico. 584 Data are shown as mean \pm S.E. (n = 4 individuals per site). Different letters indicate significant differences (p <585 0.05). 586 587 Figure 4. Carbon to nitrogen ratio for the 10 most recent consecutive pseudobulbs of L. speciosa individuals from 588 two sites in Michoacán, Mexico. Data are shown as mean \pm S.E. (n = 4 individuals per site). The vertical dashed line 589 indicates when the rescue plants were transplanted to the city. 590 Figure 5. δ^{15} N for the 10 most recent consecutive pseudobulbs of L. speciosa individuals from two sites in 591 592 Michoacán, Mexico (see Fig. 1). Data are shown as mean \pm S.E. (n = 4 individuals per site). The vertical dashed line

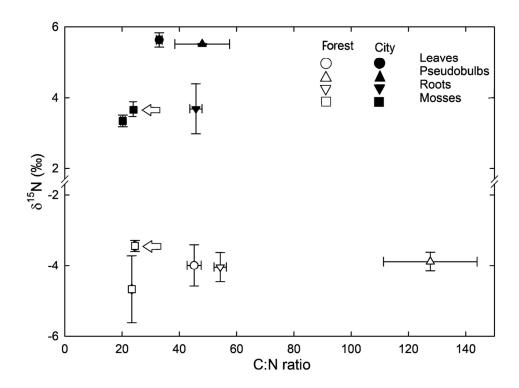
indicates the time when the rescue plants were transplanted to the city.

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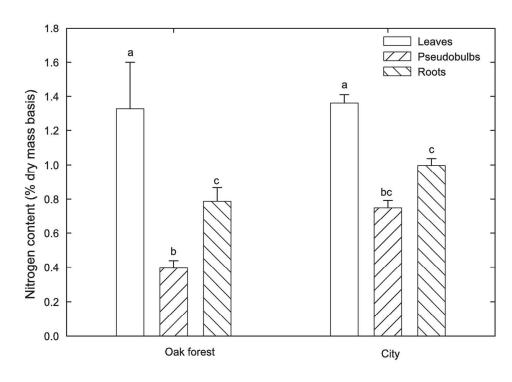
Morphology of Laelia speciosa showing the series of ten consecutive annually produced pseudobulbs. The most recent pseudobulb has been formed during the past year's growing season. The number inside each pseudobulb indicates the year it was formed.

106x68mm (300 x 300 DPI)



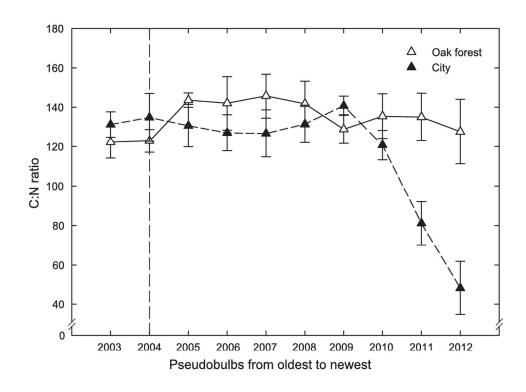
C:N ratio and $\delta 15N$ for Laelia speciosa organs and mosses growing near the orchid on the oak branches and mosses from granite or concrete from two sites in Michoacán, Mexico. Data are shown as mean \pm S.E. (n = 4 individuals per site). The arrows indicate the mosses growing on exposed bare rock at the oak forest or exposed on concrete at the city.

122x90mm (300 x 300 DPI)



Nitrogen content (dry mass basis) for the organs of L. speciosa from two sites in Michoacán, Mexico. Data are shown as mean \pm S.E. (n = 4 individuals per site). Different letters indicate significant differences (p < 0.05).

133x101mm (300 x 300 DPI)



Carbon to nitrogen ratio for the 10 most recent consecutive pseudobulbs of L. speciosa individuals from two sites in Michoacán, Mexico. Data are shown as mean \pm S.E. (n = 4 individuals per site). The vertical dashed line indicates when the rescue plants were transplanted to the city.

147x123mm (300 x 300 DPI)



